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Using dissolved noble gas and isotopic tracers to evaluate the vulnerability of groundwater resources in a small, high elevation catchment to predicted climate changes

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1. Abstract

We use noble gas concentrations and multiple isotopic tracers in groundwater and stream water in a small high elevation catchment to provide a snapshot of temperature, altitude, and physical processes at the time of recharge; and to determine subsurface residence times of different groundwater components. We identify three sources that contribute to groundwater flow: 1) seasonal groundwater recharge with short travel times, 2) water from bedrock aquifers that have elevated radiogenic ^4He , and 3) upwelling of deep fluids that have “mantle” helium and hydrothermal carbon isotope signatures. Although a bimodal distribution in apparent groundwater age indicates that groundwater storage times range from less than a year to several decades, water that recharges seasonally is the largest likely contributor to stream baseflow. Under climate change scenarios with earlier snowmelt, the groundwater that moves through the alluvial aquifer seasonally will be depleted earlier, providing less baseflow and possible extreme low flows in the creek during summer and fall. Dissolved noble gas measurements indicate recharge temperatures are 5 to 11 degrees higher than would be expected for direct influx of snowmelt, and that excess air concentrations are lower than would be expected for recharge through bedrock fractures. Instead, recharge likely occurs over diffuse vegetated areas, as indicated by $\delta^{13}\text{C}$ -DIC values that are consistent with incorporation of CO_2 from soil respiration. Recharge temperatures are close to or slightly higher than mean annual air temperature, and are consistent with recharge during May and June, when snowpack melting occurs.

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25 **2. Introduction**

26 Predicted changes in the climate will have profound impacts on water resources and
27 water management. Future climate changes in the western United States are likely
28 to include a decrease in the percentage of precipitation that falls as snow, earlier
29 onset of snow-pack melting, an increase in number of rain on snow events, and
30 changes in humidity, air temperature and soil moisture [*Dettinger and Cayan, 1995;*
31 *Howat and Tulaczyk, 2005; Maurer and Duffy, 2005; Melack, et al., 1997*]. Snowmelt
32 is an important component of groundwater recharge in high elevation watersheds
33 of the western United States e.g., [*Earman, et al., 2006*]. In these watersheds, the
34 predicted climate change impacts on snowmelt will likely alter the amount and
35 timing of groundwater recharge, which may lead to reduced groundwater
36 production, declining water tables, and reduced baseflow to streams.

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38 Groundwater aquifers in alpine and sub-alpine basins play a critical role by storing
39 and releasing snowmelt as baseflow to streams long after seasonal precipitation and
40 the disappearance of the snow pack, and in this manner significantly impact stream
41 flow and water temperature. Furthermore, geochemical hydrograph separations
42 have shown that groundwater may supply a majority of alpine streamflow during
43 peak snowmelt conditions [*Liu, et al., 2004*]. Mountain-block aquifers can also
44 provide significant recharge to mountain-front and basin-fill aquifers [*Manning and*
45 *Solomon, 2003; Manning and Solomon, 2005*]. Despite being an important part of the
46 water supply system, the recharge mechanisms, storage capacity and residence
47 times of high elevation groundwater aquifers are poorly understood and are not
48 adequately characterized by traditional methods. The net change in recharge to
49 mountain aquifers due to alterations in the timing of snowpack melting is not
50 known in sign or magnitude, making it difficult to predict the response of these
51 hydrological systems to climate change.

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53 Dissolved gases are especially well suited to examining recharge processes and
54 groundwater transport in alpine basins, with their large gradients in altitude and

temperature. Dissolved noble gases provide a snapshot of recharge water temperature and physical processes at the time of recharge. Dissolved inorganic carbon concentration, in combination with carbon isotope compositions, are useful for delineating the location of recharge and mixing of different water sources. When combined with measurements of tritium, helium isotopes provide a means of quantifying apparent groundwater subsurface residence time, or groundwater age, over a time scale relevant to the interaction between shallow groundwater and streamflow.

Manning and Caine [2007] used dissolved gas analyses to characterize groundwater recharge and residence times in a high elevation (3300-3900m asl) alpine watershed in Colorado. They determined that permeability decreases with depth and that aquifer parameters are relatively uniform throughout much of the watershed, with mean residence times between 8 and 11 years. Plummer et al., [2001] use dissolved gases to examine groundwater residence in a mountainous region in Shenandoah National Park, Virginia. They found the shallow groundwater system to be dominated by young (<3 yr) water and observed seasonally-varying recharge temperatures, indicating shallow, seasonal recharge. In a study over a much larger geographic area, [Manning and Solomon, 2005] used dissolved noble gas results to examine mountain block recharge and subsurface flow to an adjacent basin. Rademacher et al., [2001] found that the apparent groundwater ages of springs in the Sagehen Basin, a small catchment approximately 27 km north of our study site, ranged from 1 to 36 years, and based on the chemical evolution of spring and creek waters, inferred that baseflow to the local creek was dominated by moderately old groundwater [Rademacher, et al., 2005]. Also in the Sagehen Basin, Blumhagen and Clark [2008] used carbon isotope compositions to show that dissolved inorganic carbon in spring waters was inherited from respiration of CO₂ in the soil zone. In this paper, the power of dissolved gas data to evaluate the vulnerability of water resources in high elevation, snow-dominated watersheds, is demonstrated in a small basin likely to experience altered runoff and recharge under warmer climate scenarios. Specifically, the questions that are addressed

using dissolved gas and isotopic analyses are: What are the elevation, temperature, and time period over which recharge takes place? Does water infiltrate through a soil layer/unsaturated zone or through fractures? What is the range in saturated zone residence time for the bulk of the groundwater and for the groundwater most likely to contribute to stream baseflow? One of the challenges in mountain hydrology is access to adequate groundwater sampling points. The research reported here takes advantage of the numerous monitoring wells and production wells in the Olympic Valley groundwater basin.

3. Study Site

The Olympic Valley catchment is located 150 km east of Sacramento, California, near Lake Tahoe in the Sierra Nevada, and has an area of approximately 22 km² including alpine and subalpine zones. An alluvial aquifer extends eastward 4 km from the base of Granite Chief, a 2750 m peak that forms the center of Squaw Valley ski area, to the northward-flowing Truckee River (Figure 1). The valley is drained by Squaw Creek, which is formed at the confluence of two major tributaries at the west margin of the basin at elevation 1950 m, entering the Truckee River at 1850 m. The groundwater basin is underlain by Cretaceous granites of the Sierra Nevada batholith, Jurassic metasediments, and Pliocene volcanics that also form the surrounding peaks. Glacial, lacustrine, and fluvial sediments fill the valley to a maximum thickness of 55 m near the center of the 0.8 km wide valley [Gasch&Associates, 1973]. A terminal moraine at the eastern end of the basin acted as a sediment dam throughout the Quaternary period.

The hydrogeology of the valley has been examined through drill core logs and surface exposures [Hydrometrics-LLC, 2007; Kleinfelder&Associates, 1987; West-Yost&Associates, 2005] The unconsolidated valley-fill sediments act as an unconfined aquifer except where laterally discontinuous fine grained lacustrine deposits create semi-confined conditions. Three hydrostratigraphic units are loosely defined: a shallow unit consisting of fine grained lake sediments and stream deposits, a middle unit of glacial sands and gravels, and a deep unit comprising fine

grained glacial lake sediments. Coarse-grained materials predominate in the western, upstream portion of the basin and are highly permeable. Sediments become less permeable in the downstream portion of the basin, and at the terminal moraine groundwater occurrence is minimal. Groundwater also occurs in the crystalline rocks, with fractures providing secondary permeability. Four faults are mapped across the valley [NBMG, 2000], one of which is coincident with a spring ("Upwelling" spring) near monitoring wells 304/305.

Precipitation in Olympic Valley occurs mainly in the form of snow in the winter months, with a smaller amount of precipitation occurring as rain during spring and summer. A US Department of Agriculture SNOTEL site (site # 784) located at 2447 m elevation in the catchment recorded an average annual precipitation of 1684 mm for the 1982 to 2008 water years (<http://www.wcc.nrcs.usda.gov/snotel/>). Estimated isohyets [Di Luzio, et al., 2008] show mean annual precipitation increasing from 1016 mm in the east to 1650 mm in higher elevations to the west. Olympic Valley is uniquely suited to a study of alpine and sub-alpine groundwater because of its relatively simple geometry and because many wells are available for sampling. However, the natural hydrologic cycle is altered through groundwater pumping and possibly because of channelization of the Squaw Creek streambed. Groundwater is extracted to supply the needs of valley residents and businesses (approx. 6.2×10^5 m³/yr), additional resorts, private residences and the ski area (extraction unknown), snow-making (8.5×10^4 m³/yr), and for irrigation of a golf course that fills a portion of the meadow surrounding Squaw Creek (2.5×10^5 m³/yr) [Hydrometrics-LLC, 2007]. Peak water demand occurs from July to October and is about twice the wintertime demand. The main production wells are located in a cluster in the western portion of the basin (Figure 1). Although the total groundwater discharge due to pumping is a small fraction of the annual precipitation that falls in the watershed (<4%), the cluster of production wells in the upstream portion of the alluvium does capture water that would contribute to down-gradient flow or to baseflow in Squaw Creek. Fifteen monitoring well pairs are located on and around a golf course in the meadow that covers the lower valley.

A small number of horizontal wells drilled into bedrock produce water at about 0.1 L/min, which is about 10% of the flow rate of the production wells in valley alluvium.

Interaction between the groundwater production wells and stream flow is of concern because of the potential for adversely affecting habitat of brown trout and other fauna. Stream gauging upstream and downstream of the main pumping area indicates that the creek is gaining throughout the annual period of substantial stream flow (November-June). During the summer months, when stream flow decreases to $<0.05 \text{ m}^3/\text{s}$, the groundwater elevation recorded in production wells adjacent to the stream falls below the elevation of the creek bed (<http://www.svpsd.org/scada/aquiferwebdata.html>). During the year in which the study was carried out, the upper reach of the creek became dry by late summer, but deep pools and low flow persisted in the lower reach.

4. Methods

This study includes results from eight production wells, including two horizontal wells located about 100 m above the valley floor, 22 monitoring wells, and five stream sampling sites (Figure 1). Samples for DIC were passed through a 0.45 μm filter and stored in 40 mL dark glass vials with no headspace. The DIC samples were kept cold in the field and stored in a refrigerator until analysis. Tritium samples were collected in 1 L glass containers with plastic caps. Dissolved gas samples were collected using clear Tygon tubing to connect the sample vessel (8 mm inner diameter copper tubing, 250 mm long) to the wellhead of operating production wells or monitoring wells pumped by a Grundfos submersible pump. Water flowed for several minutes to purge air bubbles. The copper tubing was tapped lightly to dislodge bubbles and a visual inspection for bubbles was made. Steel clamps pinched the copper tubing flat in two locations to secure the water sample.

All analyses were performed at Lawrence Livermore National Laboratory (LLNL). Dissolved inorganic carbon (DIC) and its carbon isotope composition were determined

using the automated DIC-DOC-IRMS technique [St-Jean, 2003] consisting of an OI Analytical Model 1030 Carbon analyzer and a Micromass (now Isoprime Ltd) IsoPrime isotope ratio mass spectrometer. Carbon isotope compositions ($^{13}\text{C}/^{12}\text{C}$) are reported as delta values in per mil relative to the Vienna Pee Dee Belemnite (VPDB) reference, with an analytical uncertainty of $\pm 0.3\%$. Copper tube samples were mounted on a multi-port gas handling manifold under vacuum. Reactive gases were removed with multiple reactive metal getters. Known quantities of isotopically enriched ^{22}Ne , ^{86}Kr and ^{136}Xe were added to provide internal standards. The isotope dilution protocol used for measuring noble gas concentrations is insensitive to potential isotopic composition variation in dissolved gases (especially Ne) due to diffusive gas exchange. Noble gases were separated from one another using cryogenic adsorption. Helium was analyzed using a VG-5400 noble gas mass spectrometer. Other noble gas isotopic compositions were measured using a quadrupole mass spectrometer. The Ar abundance was determined by measuring the total noble gas sample pressure using a high-sensitivity capacitive manometer. The procedure was calibrated using water samples equilibrated with the atmosphere at a known temperature and air standards spiked with known quantities of the noble gases. Tritium concentrations were determined on 500 g sub-samples by the ^3He in-growth method (approximately 25 day accumulation time). Analytical uncertainties are approximately 1% for $^3\text{He}/^4\text{He}$, 2% for He, Ne, and Ar, and 3% for Kr and Xe. A detailed description of the data reduction routine is reported in Ekwurzel [2004].

5. Results and Discussion

A total of 34 samples from 25 wells were analyzed for noble gas, and 46 samples from wells and surface water were analyzed for tritium (Table 1). Samples were collected between April and September of 2008. A total of 60 DIC samples were collected from 30 wells, one spring, and 5 stream sampling sites (Table 1).

5.1. Excess Air

The concentration of dissolved noble gases in groundwater is virtually always greater than equilibrium solubility. The portion of gas in excess of equilibrium solubility is termed ‘excess air’ because of its compositional similarity to air

[Aeschbach-Hertig, et al., 2000; Holocher, et al., 2002]. During transport through the unsaturated zone, infiltrating water may entrain or trap air bubbles that subsequently dissolve in groundwater. Air bubbles may also become trapped in groundwater during fluctuations in the water table. The concentration of excess air provides unique information about the recharge process, including the degree to which infiltrating water incorporates unsaturated zone gas. For dissolved noble gases, addition of excess air has the greatest relative impact on He and Ne concentrations because the equilibrium components of these gases are relatively small.

A common way to represent the amount of excess air is as percent excess Ne, or ΔNe (excess Ne relative to equilibrium component; Table 2). Neon concentrations are used in determining excess air because Ne can be assumed to derive solely from the atmosphere and because Ne is measured with high precision. Excess air may be fractionated during the recharge process and Aeschbach-Hertig, et al. [1999], and later Cey et al., [2008] examined optimization models to treat fractionated excess air and calculate noble gas recharge temperatures. For samples from Olympic Valley, it was not necessary to use fractionation models because all but two samples (both from MW-PJOW) showed an unfractionated pattern in gas concentrations and a good fit with an unfractionated model of the dissolved noble gas composition. As discussed below, MW-PJOW is located upstream of the other wells in the study area, and unlike the other wells, is recharged mainly from the creek during periods of high flow. The observed fractionation of excess air in MW-PJOW, whereby heavier gases are retained relative to lighter gases (as evidenced by a relatively high ratio of Ne to Xe), is likely the result of relatively large fluctuations in the water table and poor gas confinement.

Recently, Cey et al. [2008] interpreted over 900 analyses of excess air in groundwater samples from the major groundwater basins in California. Results for over 400 samples not affected by artificial recharge (which can greatly increase

excess air) are shown in Figure 2, compared with excess air results for 35 samples from Olympic Valley and 31 wells from Handcart Gulch, an alpine watershed in Colorado [Manning and Caine, 2007]. Excess air in alpine basins is relatively unexplored, but Manning and Caine [2007] find high excess air concentrations, as shown on Figure 2, associated with bedrock wells, and attribute the high values to large fluctuations in the water table expected for recharge through fractures. Similarly, Plummer et al., [2001] found higher concentrations of excess air in groundwater from wells in fractured rock compared to other wells in Shenandoah National Park. Very low excess air concentrations are expected where an unsaturated zone is not present, as for continuous stream recharge where the water table intersects the surface [Beyerle, et al., 1999; Kipfer, et al., 2002].

Excess air concentrations in Olympic Valley wells range from ΔN_2 of 16% to 75%, with an average value of 34% and a median value of 31%, in line with the median value of 28% for the large California data set. These values argue against substantial recharge through fractures. The observed range for Olympic Valley samples suggests that groundwater recharges through an unsaturated zone, but not under conditions of high hydraulic head, large fluctuations in water table height, or very high infiltration rates. Interestingly, this observation holds true for horizontal well samples, which are drilled into fractured bedrock, indicating that recharge through soils overlying the bedrock is likely even at higher elevation in this basin. Average excess air concentrations also argue against stream recharge as the predominant recharge mechanism.

5.2. Recharge Temperature

Solubilities of the noble gases in fresh water vary as a function of temperature and pressure and are well known from theoretical and empirical studies [Andrews, 1992]. The strong temperature dependence, especially for the heavy gases, makes it possible to determine the temperature at the water table at the time of recharge [Stute and Schlosser, 1999]. Cey et al., [2009] demonstrate that recharge

temperatures determined from noble gas concentrations are very close to measured water table temperatures. Using inverse modeling, the recharge temperature and pressure are determined from noble gas concentrations. However, simultaneous estimation of both temperature and pressure results in a poorly resolved solution because these two parameters are strongly correlated. In general, pressure (or elevation) is more easily constrained by geographic conditions; once an elevation is estimated, temperature is well constrained.

Noble gas recharge temperatures (NGRTs) are calculated from Xe concentrations, after subtraction of the excess air component. Ar and Kr concentrations are used to check the goodness of fit between measured and modeled values, which is reported as X^2 (Table 2). For this sample set, X^2 values are all <11 , with 27 out of 35 samples <2 , and an average value of 1.5, indicating an acceptable fit for all samples. NGRTs, calculated assuming a recharge elevation of 1950 m, range from 5.3 ± 0.6 to 11.4 ± 0.8 °C, with an average value of 7.7 °C. The assumed recharge elevation of 1950 m is close to the break in slope between the surrounding mountainous area and the valley floor, and about 50 m above wellhead elevations, which is a likely elevation of recharge for the basin groundwaters.

Increasing the elevation in the recharge calculations will lead to lower calculated NGRTs. We can therefore calculate a minimum recharge temperature by assuming that the sample was recharged at the very top of the catchment (2750 MASL).

Assuming the maximum recharge elevation for the catchment results in a minimum recharge temperature of 3.0 ± 0.6 °C. The highest recharge temperature, assuming maximum elevation, is 8.9 ± 0.7 , and the mean of all the samples is lowered to 5.4 °C.

It is unlikely that any significant proportion of the groundwaters are recharged at this high elevation given the very small area available. A wider range of recharge elevations that covers much of the surface area, from 2300 m to 1950 m, gives a mean recharge temperature that ranges from 6.7 °C to 7.7 °C. Given that this difference is close to the analytical uncertainty for most measurements, we will use

1950 m as an estimate for elevation when calculating recharge temperatures for most wells. We will use a recharge elevation of 2050 m for the horizontal wells, since they are located approximately 100 m above the valley floor.

Even the minimum possible recharge temperatures calculated for the Olympic Valley wells are significantly higher than the melting point of snow and ice. If we assume that most recharge originates as snowmelt, then the water temperature must have increased prior to reaching the water table. This observation would indicate that an unsaturated zone is present during recharge rather than a direct connection between groundwater and the land surface, and that the infiltrating water has a residence time that allows for equilibration with the soil gas temperature.

In general, soil temperatures near the surface show a damped version of surface temperature variations, but deeper in the unsaturated zone, temperatures approach the mean annual air temperature (MAAT) e.g., [Cey, *et al.*, 2009; Kipfer, *et al.*, 2002]. Flint *et al.* [2008] measured soil temperatures at a site near Yosemite at a similar elevation (2130 masl) to the Olympic Valley, and found that soil temperatures were fairly stable and increased with depth between 10 and 72 cm under a melting snowpack. Once the snowpack disappeared, Flint *et al.* [2008] observed a rapid increase in soil temperatures.

The recharge temperatures calculated for Olympic Valley wells fall close to or slightly above the long-term MAATs reported for the nearby Tahoe City NASA GISS Climate Station (http://data.giss.nasa.gov/gistemp/station_data/), which is located approximately 7 km to the southeast of Olympic Valley at an elevation of 1899 m (Figure 3). In most cases soil gas temperatures are slightly greater (1 to 3°C) than the MAAT, so it is common for NGRTs to be slightly greater than MAAT [Kipfer, *et al.*, 2002]. Groundwaters with recharge temperatures more than 2°C above the MAAT may have recharged during warmer months, after most of the snowpack had melted. The average maximum recharge temperature for groundwater samples (7.7 °C) falls

very close to the long-term average air temperature for May (7.8 °C), when there is significant snowmelt (Figure 3C). The higher recharge temperatures fall closer to the long-term mean air temperature in June (11.4 °C). This overlap between recharge temperatures and monthly mean air temperatures suggests that most recharge takes place over a 2-3 month period during the snowmelt season. There is no discernable trend toward higher or lower recharge temperature values with apparent age [Figure 3B]. However, the large range in NGRTs observed in samples with mean apparent ages ≤ 1 year (discussed in the next section) likely shows the effects of seasonal recharge, with lower temperatures reflecting recharge at the beginning of the melting period and higher temperatures reflecting late season recharge. In wells producing water of mixed age, NGRTs represent mean, integrated values for the mixtures.

5.3. Tritium concentrations and Groundwater ages

All but two of the samples have tritium (^3H ; half life 12.32 yr) concentrations above the detection limit of 1 pCi/L (Table 1), signaling the presence of groundwater recharged within the last 50 years. Most of the tritium concentrations overlap with the expected range for modern day precipitation, which limits the utility of tritium concentrations alone for determining ages. This range also overlaps with the tritium concentration measured in Squaw Creek, which ranged from 9.7 pCi/L in September, 2008 to 12.4 pCi/L in April, 2008, with a mean for six measurements of 11.1 pCi/L (Table 1). Well values at the high end of the observed range are the result of a contribution from global fallout from nuclear testing, while observed low values are the result of decay, or dilution with older, tritium-free water.

Mean apparent groundwater ages, calculated from tritium and tritiogenic ^3He concentrations, are shown in Table 2. In order to determine tritiogenic ^3He , the measured ^3He and ^4He must be adjusted for contributions from the atmosphere (equilibrium solubility and excess air) and from subsurface sources [Cook and Solomon, 1997; Ekwurzel, et al., 1994; Schlosser, et al., 1988; Schlosser, et al., 1989]. A significant buildup of radiogenic ^4He due to decay of U and Th in crustal rocks

takes place as the saturated zone residence time increases [Andrews, et al., 1985; Torgersen and Clarke, 1985]. Radiogenic ^4He is the portion of measured ^4He remaining after subtracting solubility and excess air components. In addition, magmatic fluids can contribute dissolved helium that has a much higher $^3\text{He}/^4\text{He}$ ratio than atmospheric or crustal sources. The observed tritium and dissolved helium compositions indicate that all of these components must be considered in the analysis of Olympic Valley groundwater.

In examining samples with a mixture of ages, it is useful to determine the fraction of the mixture that recharged before about 1950 (the time of large increases in global atmospheric ^3H due to nuclear weapons testing). The reported tritium-helium age is the mean apparent age of the portion of the sample that contains tritium above the detection limit. A rough estimate of the 'percent pre-modern' is determined by comparing the initial ^3H in a sample (i.e., measured $^3\text{H} + ^3\text{He}_{\text{tritogenic}}$) with the ^3H in precipitation at the time and location of recharge (Figure 4). The nearest IAEA GNIP stations where ^3H in precipitation data were collected are Santa Maria, CA, and Menlo Park, CA (http://www-naweb.iaea.org/naweb/ih/GNIP/IHS_GNIP.html), but these are incomplete records in very different physiographic settings than Olympic Valley. Figure 4 shows an exponential fit to mean annual averaged data for Western North America, along with measured values from two cities with long records. Several points for groundwater samples from Olympic Valley with apparent ages older than 10 years fall below the curve. If one assumes 'piston flow' of groundwater, without mixing of post-modern groundwater during transport, the percentage of pre-modern water is calculated according to the difference between the expected value on the curve, and the observed value below the curve. For exponential mixing of post-modern water [Cook and Böhlke, 1999], the initial ^3H curve is generally somewhat lower than the smoothed curve, so smaller pre-modern percentages would be calculated. In any case, samples with the oldest tritium-helium ages are mixed with a significant component of pre-modern water.

Olympic Valley groundwater samples fall into three groupings: 1) samples with very

young mean apparent ages and little pre-modern water, 2) samples with a component of relatively young, tritiated water mixed with older, pre-modern water containing radiogenic ^4He , and 3) samples with a smaller component of young water, mixed with older water and magmatic fluids.

The three groups are distinguishable on a plot of $^3\text{He}/^4\text{He}$ versus Ne/He (Figure 5). On this plot, Ne and He concentrations are adjusted by subtracting the excess air component of each. For comparison, values for air saturated water at 8°C are also shown. Ne has only an atmospheric source, while He may be affected by the build-up of crustal He, accumulation of tritiogenic ^3He , or addition of magmatic He. Crustal He can contribute both ^3He (via an α, n reaction on ^6Li) and ^4He (via α decay of natural U and Th); the effect is insignificant for ^3He but can be very large for ^4He , so an increase in crustal He results in a decrease in Ne/He . Magmatic He sources have high $^3\text{He}/^4\text{He}$ and typically high He concentrations. On Figure 5, samples with magmatic He stand out as having high $^3\text{He}/^4\text{He}$ and low Ne/He , while samples with a component of radiogenic ^4He have Ne/He ratios lower than air saturated water (corresponding to wells with radiogenic ^4He concentrations $>2 \times 10^{-9} \text{ cm}^3\text{STP/gH}_2\text{O}$ in Table 2). Addition of tritiogenic ^3He due to decay of ^3H causes an increase in $^3\text{He}/^4\text{He}$ above values expected for atmospheric sources of He.

5.3.1. Recent recharge

In the first group are two monitoring wells and four production wells that yield water with a mean apparent age of less than 1 year (with a 2 sigma analytical uncertainty of about 1 year) and no detectable pre-modern water. Groundwater ages of 1 year or less in long screened, high flow wells are unusual; fewer than 2% of drinking water wells examined under California's Groundwater Ambient Monitoring and Assessment (GAMA) program (<http://www.swrcb.ca.gov/gama/>) ($n=1317$) have apparent tritium-helium ages ≤ 1 year. The wells producing very recent recharge are located in the western portion of the basin where coarse-grained

glacial and fluvial sediments prevail, which likely exhibit high hydraulic conductivity. This vigorous flow system is recharged on a short time scale (one year or less) and over a limited spatial extent (given the short time period for saturated zone transport). The production wells in this group, with screened intervals from 11 to 20 m long, do not reach depths near the bedrock basement (Figure 6). Although the creek is not likely a major source of recharge to these wells (as discussed below), the predominance of young groundwater in the alluvial aquifer suggests that it is this young component that likely provides much of the baseflow to the stream. Under climate change scenarios with earlier snowmelt and runoff, this groundwater reservoir will be depleted earlier, providing less baseflow and possible extreme low flows in the creek during summer and fall.

Several of the lower valley monitoring wells and the two horizontal wells (Figure 7) exhibit somewhat older ^3H - ^3He apparent ages. The apparent groundwater ages calculated for these wells give the groundwater age histogram its bimodal character (Figure 7). These wells are grouped with wells dominated by relatively recent recharge because they do not share dissolved gas characteristics associated with bedrock groundwater. The wells do not produce groundwater containing radiogenic ^4He , they just have higher concentrations of tritiogenic ^3He than samples with <1 year ages (Figure 5). In contrast to the production wells with <1 year mean ages, many of these wells are screened in lower permeability media, which includes near surface fine grained sediments in the lower valley and near surface fractured rock (in the case of the two horizontal wells). This older component may contribute to stream baseflow in the lower reaches of Squaw Creek, given its occurrence in shallow monitoring wells adjacent to the creek. Significant flow and deep pools are observed in downstream reaches of the creek later in the water year than in upstream reaches.

5.3.2. Bedrock aquifer flow paths

The remaining production wells (SVPSD Well 1, SVPSD Well 2), well T4, MW 5D, and five of the valley monitoring wells, fall into the second group, drawing a component

of significantly older groundwater as evidenced by a concentration of radiogenic ^4He greater than $2 \times 10^{-9} \text{ cm}^3\text{STP/g}$ and high pre-modern fractions (Table 2). These wells produce mixed aged water, as they also all contain tritium and have mean apparent groundwater ages (for the portion of the water containing tritium) of less than 50 years. All of these wells tap the deeper flow system associated with bedrock that underlies the alluvial fill, either being partially screened in bedrock or being situated near a major fault (Figure 6).

Granitic rocks have comparatively high U and Th concentrations, which can result in a relatively high radiogenic ^4He production rate [Andrews, et al., 1989]. In addition, glacial tills and weathered granites have been shown to exhibit high ^4He release rates into circulating groundwater [Beyerle, et al., 1999; Van der Hoven, et al., 2005]. Nonetheless, radiogenic ^4He concentrations in affected Olympic Valley wells are low in comparison to production wells affected by crustal He in bedrock wells elsewhere [Holocher, et al., 2001; Manning and Caine, 2007] and in deep supply wells elsewhere in California [Van der Hoven and Moran, in review]. Since this component is not observed in many of the production and monitoring wells screened exclusively in alluvium, the crustal fluid is not likely being produced within the alluvium but may be related to diffusion of ^4He from low permeability bedrock at the base of the alluvial aquifer. Although present at depth, and clearly affecting wells that directly tap bedrock groundwater or are affected by focused flow along faults (Figure 6), this component is minor in comparison to the very young groundwater component, and is not likely to play a significant role in stream interaction or baseflow to the stream.

5.3.3. Upwelling magmatic fluids

Samples that fall into group three have smaller components of recent recharge, along with an older component containing crustal He, and a component of magmatic fluid. Recently, Kulongoski et al. [2005] and Saar [2005] have presented methods for quantifying mixing proportions for groundwaters that have crustal, magmatic, and tritogenic components, as revealed by examination of the isotopic composition

of dissolved helium. Compared to samples from those studies, Olympic Valley groundwater samples have much smaller magmatic and crustal components. Olympic Valley samples are dominated by atmospheric (equilibrium solubility and excess air) and tritiogenic ^3He components. However, because the magmatic $^3\text{He}/^4\text{He}$ ratio is drastically different from the $^3\text{He}/^4\text{He}$ ratio for other helium sources, the presence of a small component of magmatic fluid in four of the samples makes determination of a ^3H - ^3He age highly unconstrained (age labeled as “ ^3He excess” in Table 2). These samples lie along a transect that lines up with an active fault, Valley Fault 3 (Figure 1 and Figure 6), and are clearly affected by fluid that emanates from a deep, magmatic source. One of these samples, MW 329, contains less than 1 pCi/L ^3H and thus is a mixture of only magmatic and crustal components. A simple linear mixing calculation for this sample (using $^3\text{He}/^4\text{He}_{\text{magmatic}} = 1.22 \times 10^{-5}$; [Graham, 2002] and $^3\text{He}/^4\text{He}_{\text{crustal}} = 6 \times 10^{-7}$ [Van der Hoven and Moran, in review]) results in estimates for the magmatic component of only 7%. The other three wells along Valley Fault 3 contain tritium, and tritiogenic ^3He , and are complex mixtures of relatively recent recharge, older water with a significant radiogenic ^4He component, and magmatic fluid that reaches shallow groundwater via the active fault. One additional monitoring well sample from the lower valley (MW302) may likewise be affected by magmatic He.

5.4. Dissolved Inorganic Carbon

Water samples from the horizontal and production wells contained 16 to 32 mg/L C as DIC with $\delta^{13}\text{C}$ -DIC values that ranged from -19.1 to -16.4 ‰. The lower valley monitoring wells have $\delta^{13}\text{C}$ -DIC values that range from -20.9 to 4.2 ‰, with concentrations from 20 to 222 mg/L C. Stream waters are low in DIC concentration (3-13 mg/L C), and have $\delta^{13}\text{C}$ values that range from -11.0 to -2.6 ‰.

The carbon isotope values for the production well groundwater samples ($\delta^{13}\text{C}$ -DIC of -19.1 to -16.4 ‰) are consistent with the incorporation of soil CO_2 during recharge e.g., [Blumhagen and Clark, 2008; Cerling, et al., 1991], which reflects a mix

of respiration CO₂ and atmospheric CO₂ sources. The incorporation of soil CO₂ in the production well groundwater suggests that recharge occurs in subalpine areas with developed soils such as the vegetated slopes surrounding the valley, as opposed to the bare rock exposures that form some of the highest elevation gradients surrounding the valley.

When plotted against 1/[DIC], the $\delta^{13}\text{C}$ -DIC values indicate mixing between three dominant sources (Figure 8): 1) groundwater recharged through the soil zone, 2) recharge from Squaw Creek, and 3) upwelling of magmatic fluids. The compositions of these end members are discussed below.

5.4.1. Recharge through the soil zone

The concentration and isotopic composition of soil CO₂ vary in relation to the respiration rate [Cerling, *et al.*, 1991]. Consequently, recharge through soils with a range of soil respiration rates might be expected to result in a range of DIC concentrations and isotopic compositions. For this reason we selected two potential end members for groundwaters recharged through the soil zone. In both cases, wells with moderate groundwater ages (11 to 16 years) were selected to avoid wells that might receive significant recharge from the creek. The average values of [DIC] (29 mg/L C) and $\delta^{13}\text{C}$ -DIC (-17.5 ‰) for the Horizontal Wells were selected as the first groundwater end member. The Horizontal wells likely see recharge through soils at higher elevations than the valley monitoring wells. Monitoring well MW-5D was selected as the second groundwater end member, and has an average [DIC] of 17.2 mg/L C and an average $\delta^{13}\text{C}$ -DIC value of -19.1 ‰.

5.4.2. Squaw Creek Water

The concentrations and isotopic compositions of DIC in the stream water samples are generally consistent with DIC derived from equilibration with atmospheric CO₂, which has a $\delta^{13}\text{C}$ value of approximately -8 ‰. Some of the stream $\delta^{13}\text{C}$ -DIC values are slightly higher than the $\delta^{13}\text{C}$ value of air, suggesting that perhaps some of the

stream DIC is derived from mineral weathering. An end member for Squaw Creek was defined by the average value of samples collected at the lower ends of the southern and northern tributaries to Squaw Creek, above the confluence (Sites 1 and 2 in Figure 1), which have an average [DIC] of 4.6 mg/L C, and an average $\delta^{13}\text{C}$ -DIC value of -4.8 ‰.

The downstream sampling sites tend to have lower $\delta^{13}\text{C}$ -DIC values and higher concentrations than the upper parts of the creek, which likely indicates an influx of groundwater along the stream channel (also indicated by stream gauge observations, as noted above). For the purposes of this study, the stream will be considered only as a possible recharge source. A more detailed study quantifying the groundwater influx to the stream is ongoing, and will not be discussed further here.

5.4.3. Upwelling magmatic water

The monitoring wells have a much broader range of DIC concentrations and isotopic compositions. Many of the monitoring wells have DIC compositions similar to the production wells. However, wells along Valley Fault #3 have much higher concentrations of DIC (50-222 mg/L C) and appear to be influenced by a carbon source with $\delta^{13}\text{C}$ values close to -5 ‰. Similar $\delta^{13}\text{C}$ values and high [DIC] have been linked to upwelling of magmatic fluids and seismic activity along faults around Mammoth Mountain in California [Sorey, *et al.*, 1998]. A contribution of magmatic water is also consistent with the $^3\text{He}/^4\text{He}$ ratios reported here. Well MW-330 was chosen as an end member to represent groundwaters that have interacted with the upwelling magmatic fluids. Well MW-330 was selected because it had the highest [DIC] of 222 mg/L C, and has no detectable tritium (<1 pCi/L).

5.4.4. Recharge from Squaw Creek

Based on the end members defined above, Squaw Creek does not appear to be a dominant source of recharge for most of the wells sampled in this study. The lone

exception is the monitoring well MW-PJOW, which was the uppermost valley well sampled. Based on samples collected in May and August, the [DIC] and $\delta^{13}\text{C}$ -DIC in MW-PJOW appears to be derived from between 50% and 70% creek water. The low recharge temperatures calculated for MW-PJOW may indicate that the creek recharge occurs during the cold, high flow conditions during snowmelt runoff. As an alternative, well MW-PJOW may receive recharge from the prominent rock escarpment to the west. Recharge through bare rock fractures would have essentially identical [DIC] and $\delta^{13}\text{C}$ -DIC values to creek water.

In general, the production wells have much higher [DIC] and much lower $\delta^{13}\text{C}$ -DIC values than Squaw Creek. This contrast between the isotopic composition and concentration of DIC in the production wells and DIC in stream water suggests that these waters have different sources and indicates that Squaw Creek is not a dominant source of recharge to the production wells. However, based on the end members defined above, samples from production wells SVPSD Well # 1, 2, and 3, fall within the range of [DIC] and $\delta^{13}\text{C}$ -DIC values that potentially indicate between 10% and 30% of produced water is from Squaw Creek. Quantifying such small contributions of the stream water is highly uncertain due to a strong dependence on the defined compositions of end members. Given the older apparent ages of SVPSD Wells 1 and 2, it is unlikely that they receive significant recharge from the creek. As discussed above, the slightly higher $\delta^{13}\text{C}$ -DIC values may also indicate the presence of water that was recharged over bare rock. Based on the mixing analysis of DIC, the majority of water at the production wells was recharged through a soil zone. This result is in agreement with the interpretation of excess air results and with stream and well hydrograph observations.

6. Summary and Conclusions

Dissolved gas tracers provide a powerful toolset to evaluate the vulnerability of high altitude watersheds to climate change impacts because they address key questions about recharge location and subsurface residence time. The extent to which

individual catchments are vulnerable to climate change will depend largely on the specifics of geology, topography, and climate. The dissolved gas toolset used for this study can be applied under a wide range of potential settings, but is especially useful in high elevation areas because of the steep gradients in precipitation and temperature and because active recharge and vigorous flow results in relatively young apparent groundwater ages, dateable by the tritium-helium method.

Young groundwaters dominate the alluvial aquifer in Olympic Valley, and likely account for most of the potential baseflow to Squaw Creek. These waters have apparent ages of less than one year and are therefore vulnerable to climate changes over short time scales. Mixed age components also need to be considered in studies of alpine and sub-alpine groundwater residence time. In this study, the bedrock aquifer underlying the valley fill contributes an older component that has accumulated radiogenic ^4He . In addition, He and C isotopes show the influence of magmatic fluids in shallow groundwater, especially in the area of an active fault. Evaluating the residence times of groundwater in these systems provides an important constraint on the potential responses of stream baseflow to climate change. The major findings with respect to groundwater residence times in Olympic Valley, i.e., that the alluvial aquifer experiences rapid flushing of seasonal recharge, and that significantly older fluids are found at the bedrock interface that underlies the alluvium, are similar to the major findings of Beyerle et al. [1999] in the Lisenthal aquifer of Switzerland and Plummer et al. [2001] in the Blue Ridge Mountains. Similar studies are needed in bigger catchments to determine whether an increase in scale may decrease the dominance of young waters.

NGRTs constrain the location and timing of recharge and correspond to air temperatures at the time of snowmelt. Recharge occurs mainly through soil zones where the waters incorporate CO_2 from respiration and recharging water is thermally equilibrated within the unsaturated zone. Recharge through fractures and recharge from the creek is less important in this catchment. Long term monitoring of recharge temperatures may provide a means to gauge watershed

response to climate changes such as an earlier onset of snowmelt and changes in mean air temperature.

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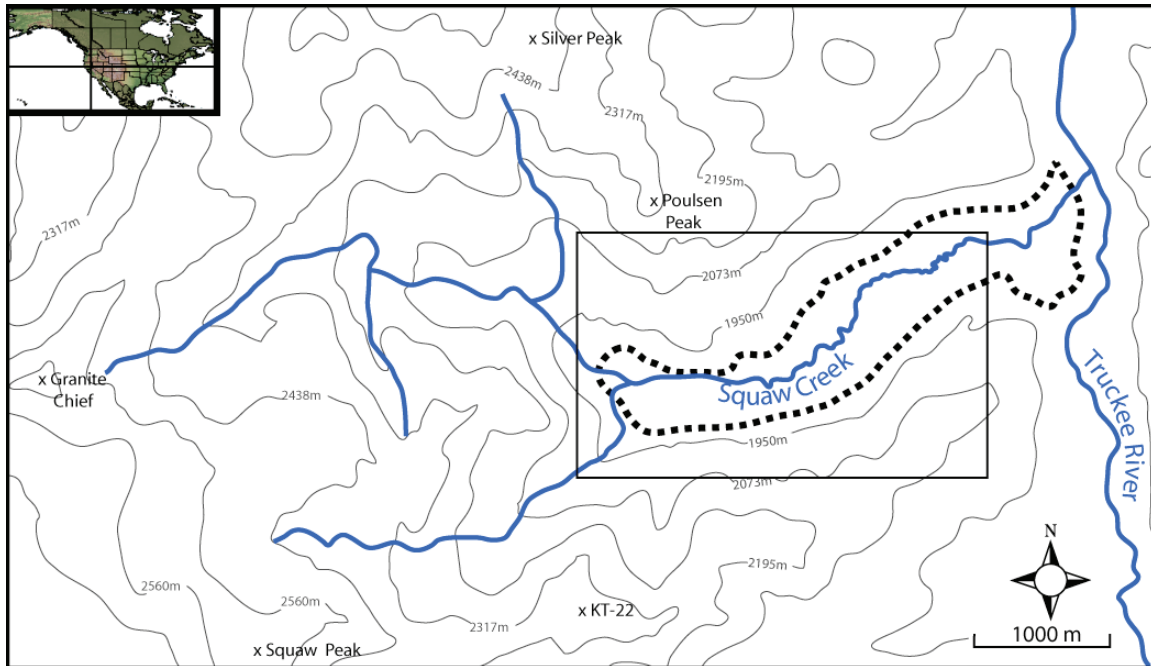
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Figures

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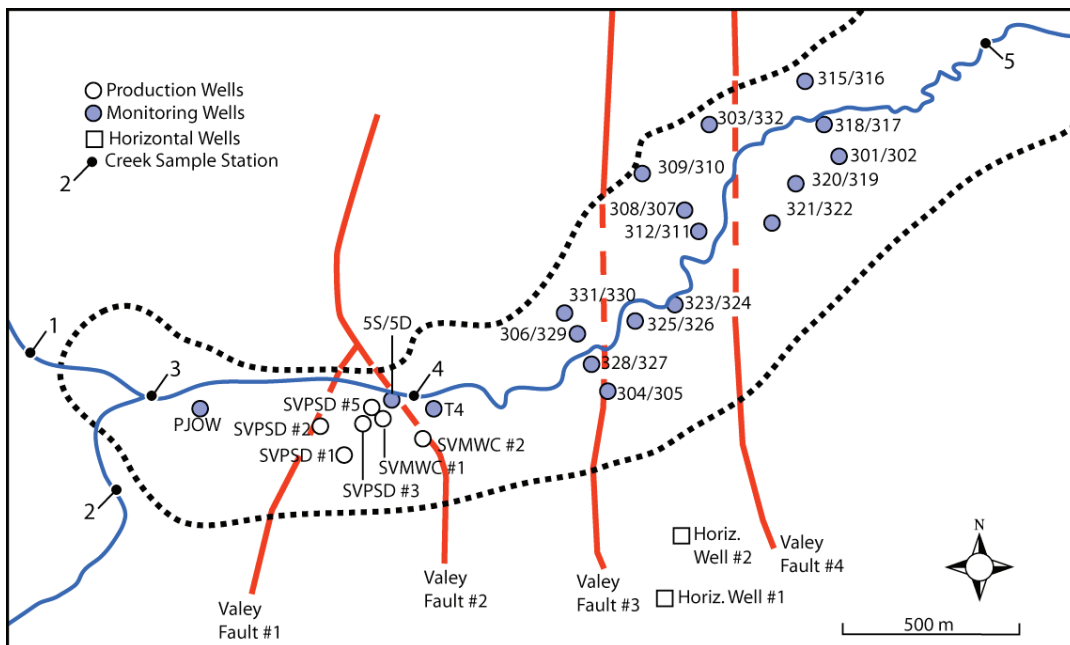


Figure 1. Topographic map (A) of the study area, and (B) locations of the wells and faults discussed in the text. The outline of the alluvial aquifer is shown as a dotted line after [Hydrometrics-LLC, 2007]. Stream sampling sites as labeled in Table 2 are 1) Shirley Canyon, 2) South Fork, 3) Confluence, 4) Trapezoid, and 5) Squaw Creek Rd Bridge.

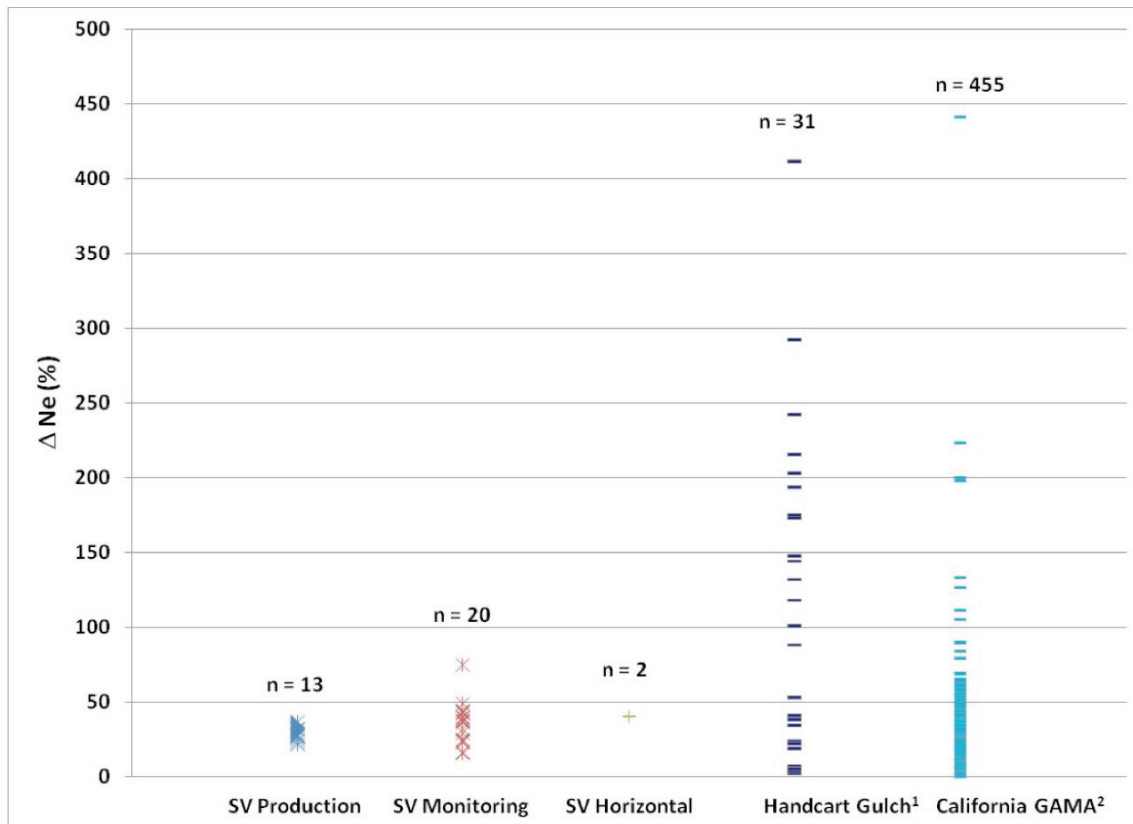


Figure 2. Observed ranges in excess air, expressed as ΔNe , for samples from Olympic Valley (categorized by well type), samples from the alpine watershed Handcart Gulch, and a large number of samples from around California sampled under the GAMA program.

¹ [Manning and Caine, 2007]

² [Cey, et al., 2008]

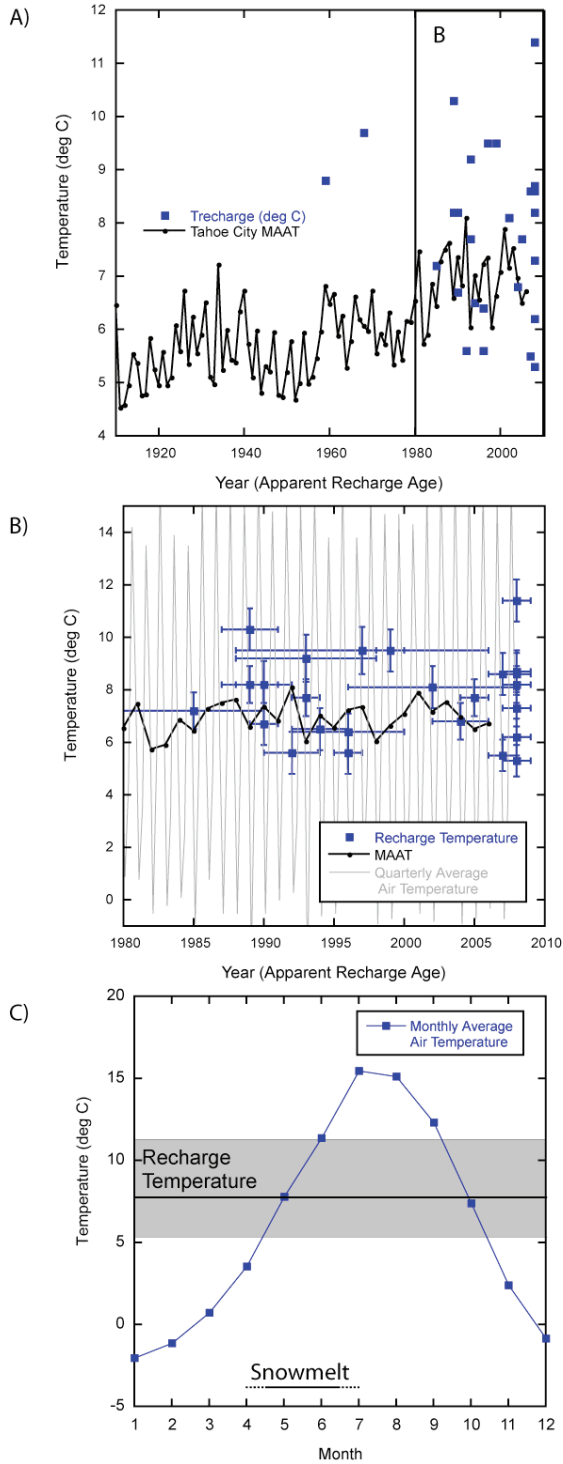


Figure 3. Noble gas recharge temperatures and apparent groundwater ages shown with mean annual air temperature (A, B), quarterly air temperature (B), and average monthly air temperature (C) at the NASA/GISS station at Tahoe City, CA. The range (shaded) and average (horizontal line) of recharge temperatures, along with the typical snowmelt season are shown on C.

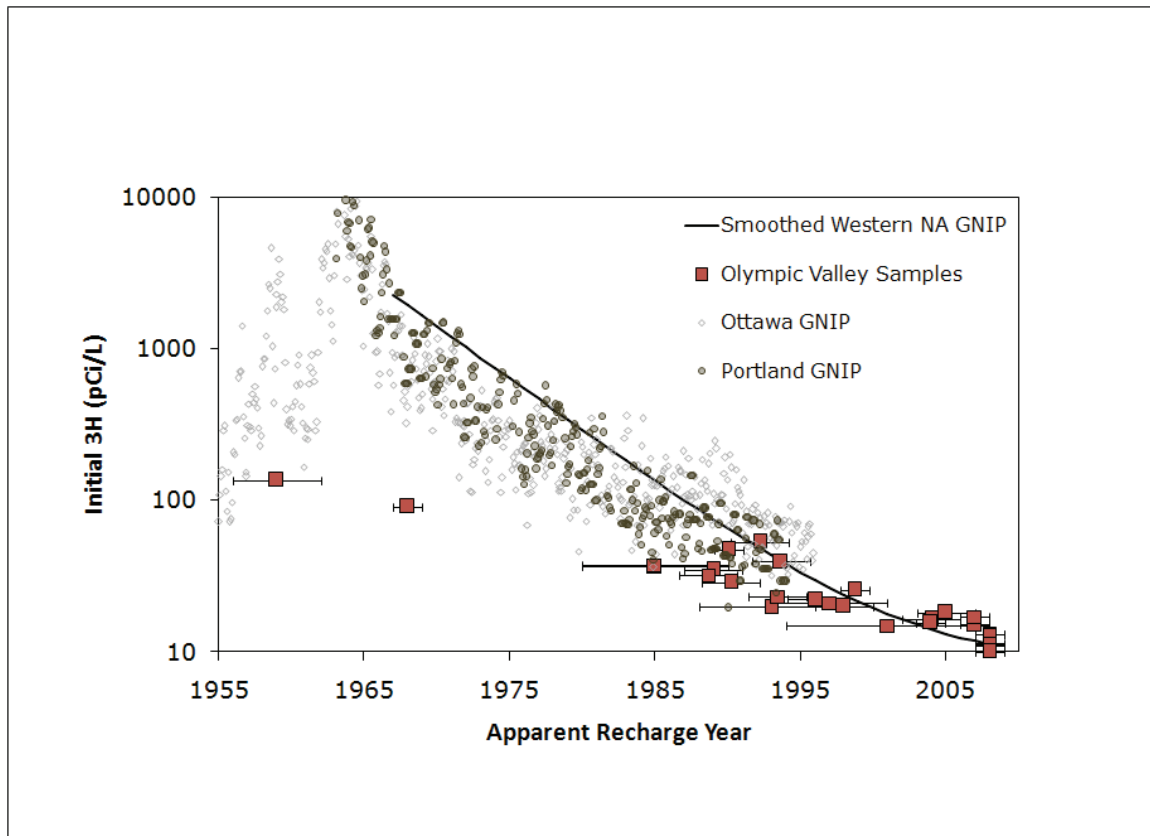


Figure 4. Tritium concentrations measured in precipitation at two locations where long IAEA GNIP records exist, along with an exponential curve that approximates mean annual values from western North America. Results for Olympic Valley well samples are plotted according to the calculated apparent tritium helium age (recharge year) and the measured tritium + tritiogenic ^3H (initial ^3H). Points that fall well below the curve contain a significant component of pre-modern water.

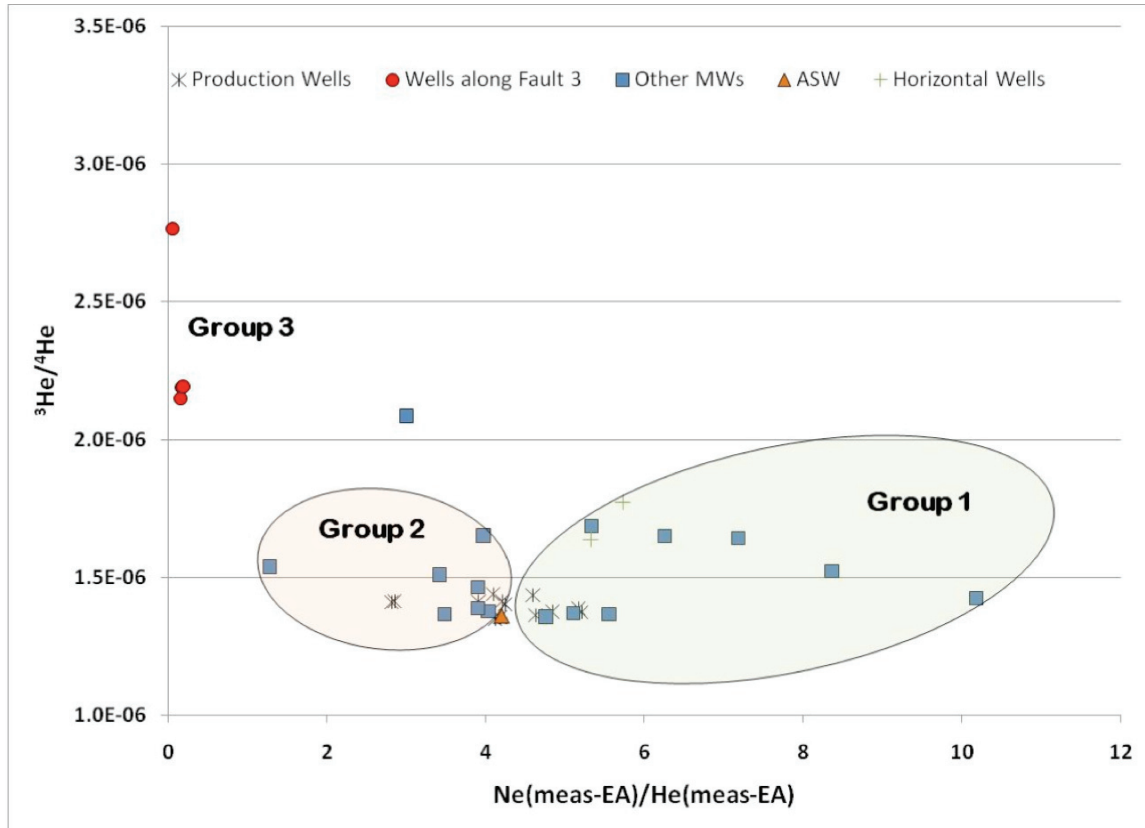


Figure 5. A plot of the ratio of Ne/He (measured concentration minus amount due to excess air) versus the measured $^3\text{He}/^4\text{He}$ ratio. Samples in Group 1 with very young ages are close to solubility values; significant amounts of tritiogenic ^3He bring some samples above the solubility ratio of 1.364×10^{-6} . Group 2 samples are affected by crustal He, which results in a decrease in Ne/He and by tritiogenic ^3He . Group 3 samples are affected by magmatic He. Orange triangle represents air saturated water at 8°C .

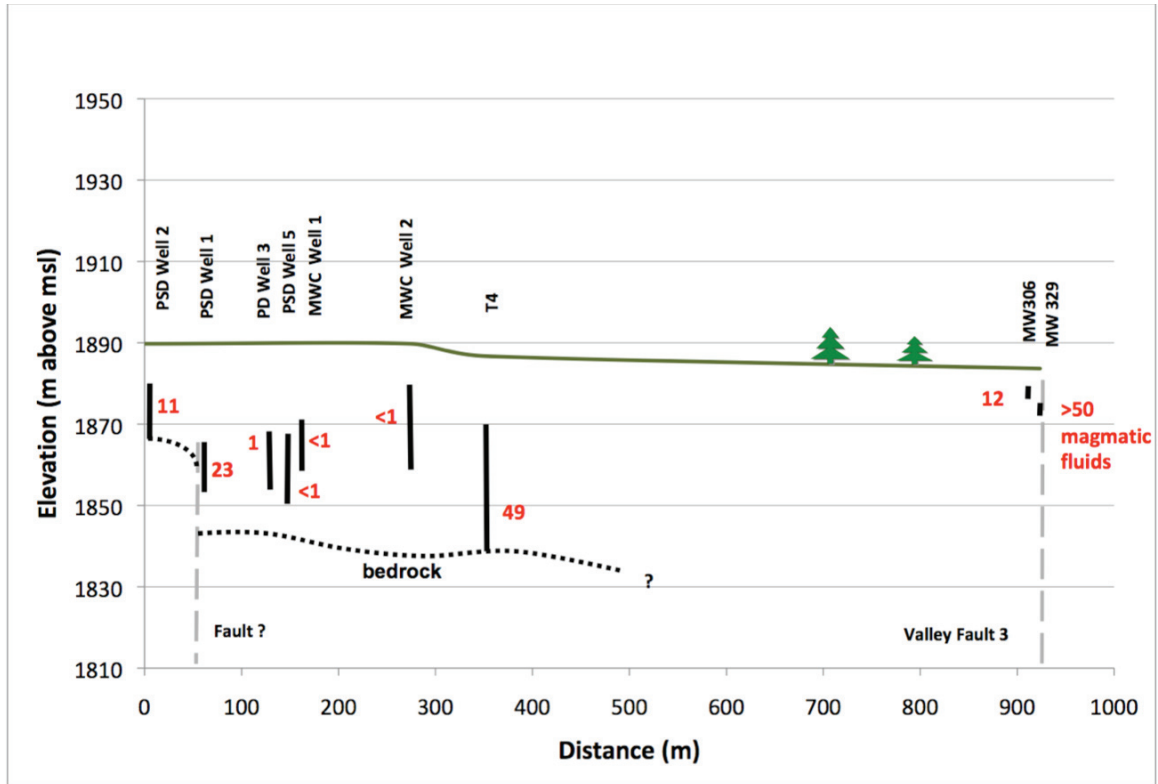


Figure 6. Schematic cross section through Olympic Valley running along Squaw Creek, showing the approximate locations of the major faults and the depth to bedrock as based on well logs [West-Yost&Associates, 2005] by seismic profiling [Gasch&Associates, 1973]. Solid vertical lines represent screened intervals for wells on or close to the cross section, with labels indicating the mean apparent tritium-helium groundwater age. Vertical exaggeration is 10X.

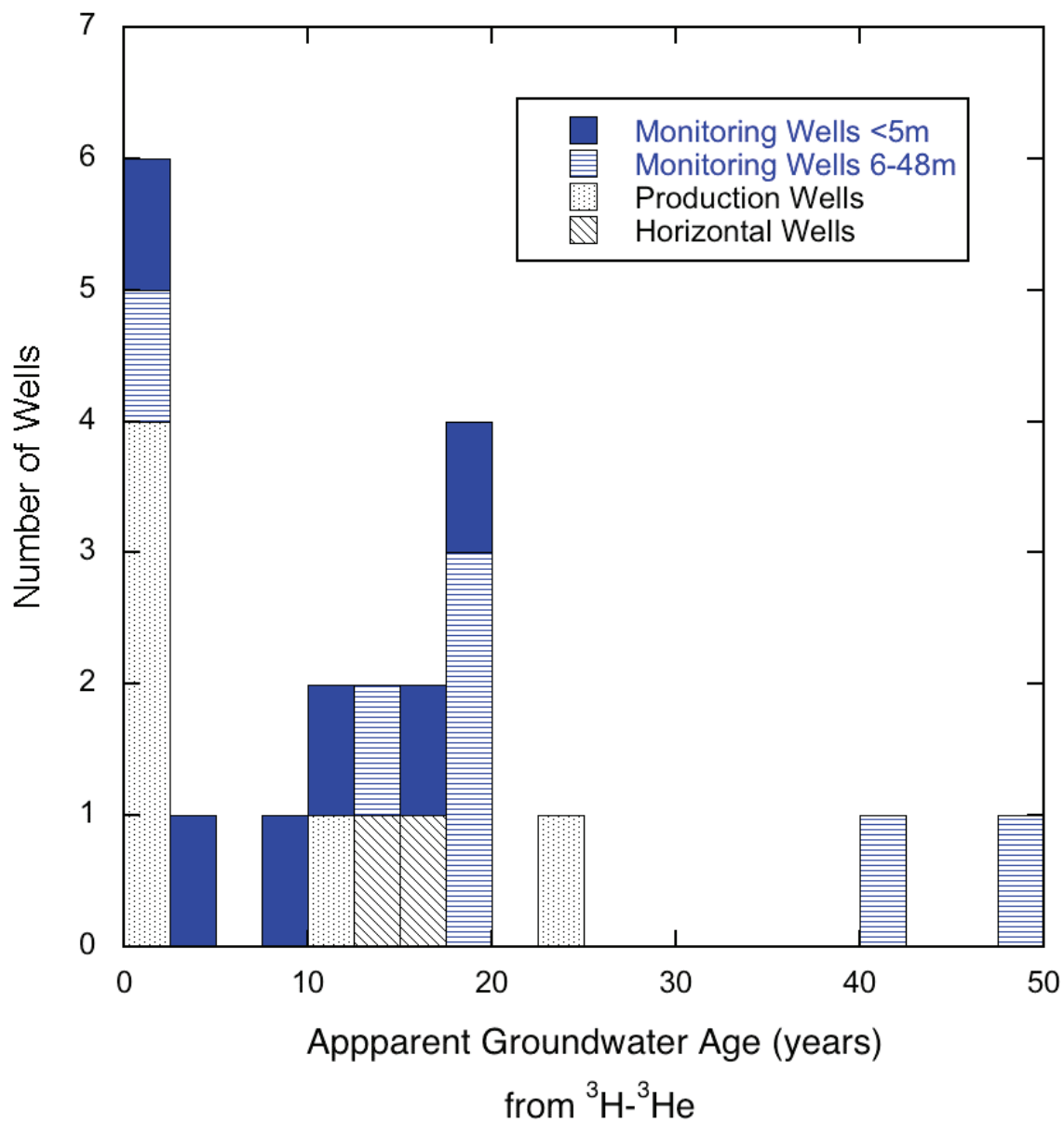


Figure 7. Histogram of apparent groundwater ages. Average apparent ages are plotted for wells that were sampled more than once.

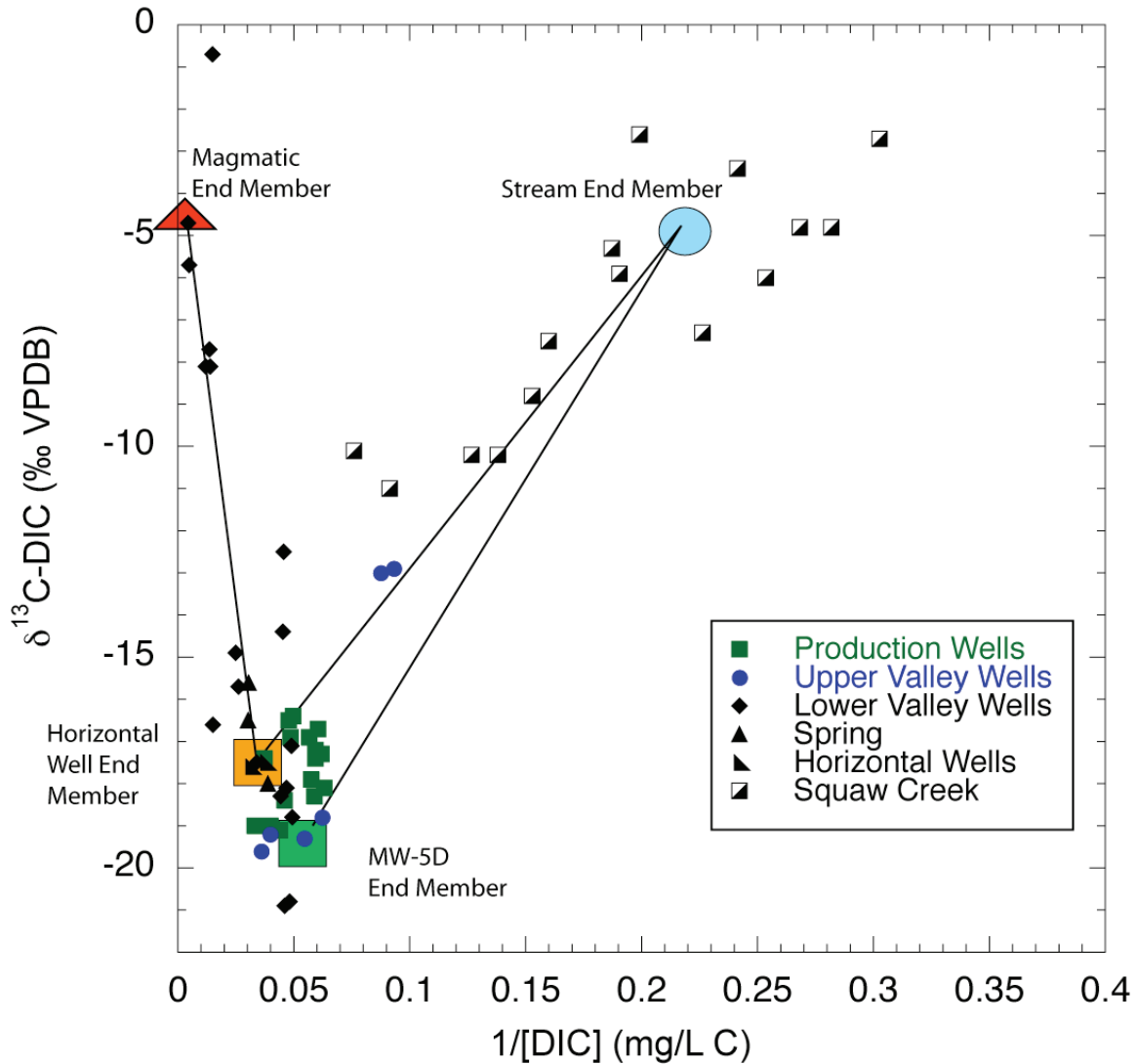


Figure 8. Stable carbon isotope compositions ($\delta^{13}\text{C-DIC}$) and inverse concentration of dissolved inorganic carbon for wells, a spring, and creek water in the Olympic Valley study area. Mixing lines are plotted between two potential end members for groundwater recharge through soils, a magmatic water end member, and a stream end member.

Table 1. Measurements of dissolved gas and isotopic compositions from horizontal wells “HW”, monitoring wells “MW”, production wells “PW”, a spring “SP”, and stream waters “SW”.

Sample Site	Collection Date	Sample Type	Screen depth (m)	DIC (mg/L C)	δ ¹³ C (‰ VPDB)	N (ppb/L)	+/-	N ₂ /He	+/-	He (cc STP/g)	+/-	Ar (cc STP/g)	+/-	Kr (cc STP/g)	+/-	Ne (cc STP/g)	+/-	Xe (cc STP/g)	+/-
Horizontal Well 1	06/19/2008	HW		26	-17.5		1.1	1.64E-06	5.09E-08	6.49E-08	2.07E-09	4.00E-04	7.99E-06	8.99E-08	2.70E-09	2.83E-07	5.67E-09	1.25E-08	3.75E-10
	06/19/2008	HW	1.2	32	-17.6	21.8	0.9	1.77E-06	5.52E-08	6.07E-08	1.94E-09	4.15E-04	8.30E-06	9.22E-08	2.77E-09	2.76E-07	5.33E-09	1.28E-08	3.85E-10
Horizontal Well 2	05/13/2008	HW	3-5	38	-15.7	10.2	0.9	1.65E-06	1.24E-08	3.56E-08	7.12E-10	3.42E-04	6.84E-06	7.75E-08	2.32E-09	1.98E-07	3.96E-09	1.15E-08	3.45E-10
	05/13/2008	MMW	12-14	40	-14.9	9.8	0.9	2.09E-06	1.56E-08	6.79E-08	1.36E-09	3.39E-04	6.78E-06	7.66E-08	2.30E-09	2.11E-07	4.22E-09	1.08E-08	3.25E-10
MMW 301	06/18/2008	MMW	3-5	20	-18.8	11.5	1.4	1.46E-06	4.25E-08	6.67E-08	1.33E-09	3.53E-04	7.07E-06	7.78E-08	2.33E-09	2.49E-07	4.99E-09	1.12E-08	3.36E-10
MMW 303	05/14/2008	MMW	3-5	72	-7.7	6.3	0.6	2.19E-06	1.64E-08	9.53E-07	1.91E-08	3.85E-04	7.70E-06	8.39E-08	2.52E-09	2.80E-07	5.60E-09	1.24E-08	3.72E-10
MMW 304	05/14/2008	MMW	11-13	73	-8.1	6.4	0.6	2.15E-06	1.61E-08	9.02E-06	2.04E-08	3.87E-04	7.73E-06	8.53E-08	2.56E-09	2.64E-07	5.28E-09	1.20E-08	3.60E-10
MMW 305	05/14/2008	MMW	2-3	21	-20.8	11.4	1.4	1.52E-06	1.14E-08	5.74E-08	1.15E-09	4.11E-04	8.21E-06	9.09E-08	2.73E-09	2.99E-07	5.97E-09	1.31E-08	3.93E-10
MMW 306	06/18/2008	MMW	3-5	22	-14.4	17.8	1.5												
MMW 309	06/18/2008	MMW	11-12	22	-20.9	10.6	1.4	1.64E-06	4.77E-08	5.30E-08	1.06E-09	3.75E-04	7.50E-06	8.19E-08	2.46E-09	2.67E-07	5.33E-09	1.18E-08	3.55E-10
MMW 310	06/18/2008	MMW	2-3	77	-4.2	11.3	1.0												
MMW 315	06/18/2008	MMW	8-10	66	-16.6	10.4	1.4												
MMW 316	06/18/2008	MMW	3-5	67	-10.7	16.9	1.1												
MMW 318	05/14/2008	MMW	17-18	29	-17.5	10.8	1.4	1.68E-06	1.26E-08	5.28E-08	1.06E-09	3.33E-04	6.67E-06	7.61E-08	2.28E-09	2.41E-07	4.81E-09	1.08E-08	3.25E-10
MMW 321	05/14/2008	MMW	3-5	21	-18.1	14.6	1.5	1.65E-06	1.24E-08	7.06E-08	1.41E-09	3.82E-04	7.64E-06	8.58E-08	2.57E-09	2.66E-07	5.33E-09	1.24E-08	3.72E-10
MMW 322	05/14/2008	MMW	11-13	83	-8.1														
MMW 327	05/14/2008	MMW	2-3	23	-18.3	12.4	1.4	1.42E-06	1.07E-08	5.37E-08	1.07E-09	3.68E-04	7.35E-06	8.67E-08	2.60E-09	2.96E-07	5.91E-09	1.25E-08	3.76E-10
MMW 328	05/14/2008	MMW	13-15	203	-5.7	0.3	0.7	2.77E-06	3.40E-08	2.91E-06	5.82E-08	4.26E-04	8.52E-06	8.95E-08	2.69E-09	3.17E-07	6.34E-09	1.32E-08	3.95E-10
MMW 329	05/14/2008	MMW	8-9	222	-4.7	-0.5	0.9												
MMW 330	05/13/2008	MMW	2-3	10	-17.1	12.0	1.0	1.51E-06	1.13E-08	6.19E-08	1.24E-09	3.50E-04	7.00E-06	8.02E-08	2.41E-09	2.13E-07	4.26E-09	1.14E-08	3.42E-10
MMW 331	05/13/2008	MMW	12-14	18	-19.3	8.7	0.6	1.36E-06	1.02E-08	7.38E-08	1.48E-09	3.63E-04	7.26E-06	7.95E-08	2.38E-09	2.58E-07	5.15E-09	1.14E-08	3.41E-10
MMW 5D	08/12/2008	MMW	12-14	16	-18.8	9.1	1.0	1.39E-06	8.05E-08	6.74E-08	5.50E-09	3.53E-04	7.05E-06	8.03E-08	2.41E-09	2.52E-07	5.03E-09	1.12E-08	3.37E-10
MMW 5S	05/13/2008	MMW	6-8	28	-19.6	11.4	0.7	1.37E-06	1.02E-08	4.84E-08	9.69E-10	3.81E-04	7.53E-06	8.58E-08	2.57E-09	2.29E-07	4.58E-09	1.16E-08	3.47E-10
MMW 5S	08/12/2008	MMW	24-26	25	-19.2	11.7	1.0	1.38E-06	7.94E-08	4.54E-08	3.70E-09	3.20E-04	6.40E-06	7.39E-08	2.22E-09	1.87E-07	4.15E-09	1.03E-08	3.08E-10
MMW 5S	05/13/2008	MMW	6-8	25	-13.0	10.3	1.0	1.37E-06	1.03E-08	4.74E-08	9.49E-10	3.43E-04	6.87E-06	8.47E-08	2.54E-09	1.87E-07	4.02E-09	1.24E-08	3.71E-10
MMW P10W	09/26/2008	MMW	24-26	11	-12.9	9.3	0.9	1.36E-06	2.09E-08	4.59E-08	3.75E-09	3.72E-04	7.39E-06	8.65E-08	2.59E-09	2.01E-07	5.27E-09	1.26E-08	3.78E-10
MMW T4	09/26/2008	MMW	16-48	22	-12.5	8.7	0.6	1.54E-06	8.79E-08	1.57E-07	3.13E-09	3.48E-04	6.97E-06	8.36E-08	2.51E-09	2.64E-07	4.74E-09	1.15E-08	3.49E-10
MMWC Well 1	08/12/2008	PW	18-30	22	-18.4	10.0	0.9	1.37E-06	1.03E-08	5.67E-08	4.63E-09	3.69E-04	7.35E-06	8.12E-08	2.44E-09	2.37E-07	4.97E-09	1.16E-08	3.42E-10
MMWC Well 2	08/12/2008	PW	11-30	23	-19.0	11.1	0.4	1.43E-06	8.32E-08	1.57E-07	1.71E-09	3.57E-04	7.75E-06	8.22E-08	2.47E-09	2.41E-07	4.61E-09	1.14E-08	3.42E-10
MMWC Well 2	08/12/2008	PW	11-30	25	-19.0	10.1	0.4	1.36E-06	1.02E-08	5.34E-08	1.71E-09	3.64E-04	7.28E-06	8.30E-08	2.49E-09	2.41E-07	4.81E-09	1.14E-08	3.42E-10
SVPSD Well 1	06/18/2008	PW	23-34	21	-16.9	9.6	2.7												
SVPSD Well 1	06/18/2008	PW	23-34	21	-16.5	10.2	2.4	1.41E-06	4.10E-08	7.50E-08	1.50E-09	3.64E-04	7.29E-06	8.29E-08	2.49E-09	2.25E-07	4.50E-09	1.19E-08	3.58E-10
SVPSD Well 1	08/12/2008	PW	23-34	20	-16.4			1.41E-06	8.18E-08	7.99E-08	6.52E-09	3.75E-04	7.50E-06	8.96E-08	2.69E-09	2.40E-07	4.80E-09	1.24E-08	3.73E-10
SVPSD Well 2	04/25/2008	PW	10-23	17	-17.2	11.3	0.5												
SVPSD Well 2	06/18/2008	PW	10-23	17	-16.7	10.0	1.0	1.44E-06	4.17E-08	5.77E-08	1.15E-09	3.65E-04	7.31E-06	8.49E-08	2.55E-09	2.26E-07	4.52E-09	1.23E-08	3.68E-10
SVPSD Well 2	08/12/2008	PW	10-23	16	-18.1			1.41E-06	8.18E-08	5.56E-08	4.54E-09	3.50E-04	7.01E-06	8.49E-08	2.55E-09	2.12E-07	4.23E-09	1.19E-08	3.67E-10
SVPSD Well 3	04/25/2008	PW	22-35	17	-17.4	10.4	0.4	1.35E-06	1.01E-08	NM		3.65E-04	7.29E-06	7.95E-08	2.39E-09	2.15E-07	4.29E-09	1.18E-08	3.54E-10
SVPSD Well 3	06/18/2008	PW	22-35	18	-16.9	10.5	1.0	1.41E-06	4.10E-08	5.43E-08	1.09E-09	3.56E-04	7.13E-06	8.29E-08	2.49E-09	2.18E-07	4.35E-09	1.20E-08	3.61E-10
SVPSD Well 5	04/25/2008	PW	17	17	-18.3	10.3	0.4	1.36E-06	1.02E-08	5.75E-08	1.15E-09	3.63E-04	7.26E-06	7.97E-08	2.39E-09	2.28E-07	4.56E-09	1.19E-08	3.56E-10
SVPSD Well 5	06/18/2008	PW	22-39	17	-17.9	16.1	2.7	1.39E-06	4.03E-08	5.88E-08	1.18E-09	3.74E-04	7.29E-06	8.21E-08	2.46E-09	2.58E-07	5.16E-09	1.16E-08	3.47E-10
SVPSD Well 5	08/12/2008	PW	22-39	16	-17.3	12.8	1.0	1.40E-06	4.37E-08	5.49E-08	1.76E-09	3.49E-04	6.98E-06	7.99E-08	2.40E-09	2.20E-07	4.41E-09	1.17E-08	3.51E-10
SVPSD Well 5	04/25/2008	SP		33	-16.5														
Upwelling	05/14/2008	SP		33	-15.6														
Upwelling	08/12/2008	SP		26	-18.0	10.7	0.9												
Upwelling	04/25/2008	SP		7	-10.2	10.5	0.4												
Confluence	06/19/2008	SW		4	-4.8														
Confluence	05/14/2008	SW		4	-3.4														
Shirley Canyon	04/25/2008	SW		5	-5.3	10.5	0.5												
Shirley Canyon	06/18/2008	SW		4	-4.8														
Shirley Canyon	06/18/2008	SW		3	-2.7														
South Fork at Bridge	04/25/2008	SW		6	-7.5	12.4	0.5												
South Fork at Bridge	05/14/2008	SW		4	-6.0														
Squaw Creek Rd bridge	06/18/2008	SW		5	-2.6														
Squaw Creek Rd bridge	04/25/2008	SW		8	-10.2	11.5	0.5												
Squaw Creek Rd bridge	05/14/2008	SW		4	-7.3														
Squaw Creek Rd bridge	06/19/2008	SW		5	-5.9														
Squaw Creek Rd bridge	08/12/2008	SW		11	-11.0														
Squaw Creek Rd bridge	09/25/2008	SW		13	-10.1	9.7	0.5												
Trapezoid	04/25/2008	SW		7	-8.8	12.0	0.5												

NM = Measurement failed

Table 2. Calculations based on noble gas concentrations and isotope ratios.

Well Name	Date	Screen depth (m)	AHe	⁴ He _{excess}	+/-	T _{radiometric} (deg C) ¹	+/-	X ₂	Age ²	+/-	% Pre-modern
Upper Valley MW's											
MW PUOW	5/13/08	24-26	17%	4.46E-09	1.63E-09	5.5	0.6	1.0	1	1	<10%
MW PUOW	8/12/08	24-26	24%	<2.0E-09		5.3	0.6	3.1	<1	1	<10%
MW 5D	5/13/08	12-14	38%	1.18E-08	4.12E-09	9.2	0.9	0.3	15	5	52%
MW 5D	8/12/08	12-14	37%	6.32E-09	3.84E-09	9.5	0.9	0.8	11	9	40%
MW 5S	5/13/08	6-8	30%	<2.0E-09		8.2	0.7	11.0	<1	1	<10%
MW 5S	8/12/08	6-8	25%	<2.0E-09		11.4	0.8	0.5	<1	1	<10%
MW T4	9/26/08	16-48	37%	9.40E-08	4.79E-09	8.8	0.9	0.4	49	3	98%
Production wells											
SVPSD Well1	6/18/08	23-34	27%	2.23E-08	2.81E-09	7.2	0.7	0.4	23	5	72%
SVPSD Well1	8/12/08	23-34	31%	2.32E-08	3.43E-09	6.2	0.7	1.2			
SVPSD Well2	6/18/08	10-23	27%	4.76E-09	2.81E-09	6.4	0.7	0.0	12	4	21%
SVPSD Well2	8/12/08	10-23	23%	5.78E-09	2.27E-09	7.1	0.7	1.3			
SVPSD Well3	4/25/08	22-35	23%	4.09E-09	2.36E-09	7.3	0.7	3.0	<1	1	<10%
SVPSD Well3	6/18/08	22-35	25%	NM		6.8	0.7	0.1	4	2	<10%
SVPSD Well5	4/25/08	22-39	28%	3.82E-09	2.90E-09	7.3	0.7	1.3	<1	1	<10%
SVPSD Well5	6/18/08	22-39	37%	<2.0E-09		8.6	0.8	0.5	1	1	<10%
SVPSD Well5	8/12/08	22-39	27%	2.66E-09	2.63E-09	7.7	0.7	0.8	3	1	<10%
MWC Well 1	4/25/08	18-30	27%	<2.0E-09		8.6	0.8	4.1	<1	1	<10%
MWC Well 1	8/12/08	18-30	32%	<2.0E-09		8.1	0.8	0.2	6	6	<10%
MWC Well 2	4/25/08	11-30	22%	<2.0E-09		6.2	0.7	7.5	<1	1	<10%
MWC Well 2	8/12/08	11-30	31%	<2.0E-09		8.7	0.8	1.4	<1	1	<10%
Horizontal Wells											
Horizontal Well 1	6/19/08		41%	<2.0E-09		6.5	0.8	0.6	14	2	<10%
Horizontal Well 2	6/19/08		40%	<2.0E-09		5.6	0.8	3.5	16	2	<10%
Lower Valley MW's											
Shallow (2-5 m BGS)											
MW 301	5/13/08	3-5	16%	<2.0E-09		7.7	0.7	0.7	15	1	43%
MW 303	6/18/08	3-5	37%	6.21E-09	3.75E-09	9.5	0.8	1.0	9	1	34%
MW 304	5/14/08	3-5	43%	8.85E-07	1.84E-08	7.0	0.9	2.0	³ He excess		NA
MW 306	5/14/08	2-3	43%	<2.0E-09		5.6	0.8	0.1	12	1	25%
MW 321	5/14/08	3-5	32%	<2.0E-09		10.3	0.8	2.6	19	2	59%
MW 328	5/14/08	3-5	44%	<2.0E-09					4	1	<10%
MW 331	5/13/08	2-3	24%	1.18E-08	2.33E-09	8.2	0.7	0.4	19	2	54%
Deep (11-18 m BGS)											
MW 302	5/13/08	12-14	24%	1.80E-08	2.29E-09	9.7	0.7	0.4	40	1	95%
MW 305	5/14/08	11-13	39%	9.54E-07	1.96E-08	7.7	0.8	0.4	³ He excess		NA
MW 310	6/18/08	11-12	40%	<2.0E-09		8.2	0.9	0.5	18	2	54%
MW 322	5/14/08	17-18	38%	7.27E-08	4.53E-09	6.7	0.8	0.3	18		52%
MW 327	5/14/08	11-13	75%	8.03E-07	2.53E-08	9.2	2.5	0.1	³ He excess	1	NA
MW 329	5/14/08	13-15	49%	2.83E-06	5.70E-08	5.9	0.9	0.8	³ He excess		NA

1 An elevation of 1950 m used to determine pressure, except at the two horizontal wells, where an elevation of 2050 m is applied.

2 A crustal ³He/⁴He ratio of 6x10⁻⁷ is used in the age determination for samples with [He_{meas} - (He_{sol} + He_{xs air})] >2x10⁻⁹.